CONSERVATION

Widespread diversity deficits of coral reef sharks and rays

Colin A. Simpfendorfer^{1,2}*, Michael R. Heithaus³, Michelle R. Heupel^{2,4}, M. Aaron MacNeil⁵, Mark Meekan⁶, Euan Harvey⁷, C. Samantha Sherman^{1,8}, Leanne M. Currey-Randall⁴, Jordan S. Goetze^{9,10}, Jeremy J. Kiszka³, Matthew J. Rees^{6,11}, Conrad W. Speed⁶, Vinay Udyawer¹², Mark E. Bond³, Kathryn I. Flowers^{3,13}, Gina M. Clementi³, Jasmine Valentin-Albanese¹⁴, M. Shiham Adam¹⁵, Khadeeja Ali^{3,16}, Jacob Asher¹⁷ Eva Aylagas¹⁷, Océane Beaufort¹⁸, Cecilie Benjamin¹⁹, Anthony T. F. Bernard^{20,21}, Michael L. Berumen²², Stacy Bierwagen⁴, Chico Birrell²³, Erika Bonnema³, Rosalind M. K. Bown²⁴, Edward J. Brooks²⁵, J. Jed Brown²⁶, Dayne Buddo²⁷, Patrick J. Burke^{28,29}, Camila Cáceres³, Marta Cambra^{30,31} Diego Cardeñosa³, Jeffrey C. Carrier³², Sara Casareto³, Jennifer E. Caselle³³, Venkatesh Charloo³⁴, Joshua E. Cinner³⁵, Thomas Claverie³⁶, Eric E. G. Clua^{37,38}, Jesse E. M. Cochran²², Neil Cook^{39,40}, Jessica E. Cramp^{41,42}, Brooke M. D'Alberto^{1,43}, Martin de Graaf⁴⁴, Mareike C. Dornhege⁴⁵, Mario Espinoza^{30,31}, Andy Estep⁴⁶, Lanya Fanovich⁴⁰, Naomi F. Farabaugh³, Daniel Fernando²⁴, Carlos E. L. Ferreira⁴⁷, Candace Y. A. Fields^{3,25}, Anna L. Flam⁴⁸, Camilla Floros^{49,50}, Virginia Fourqurean^{51,52}, Laura Gajdzik^{22,53}, Laura García Barcía³, Ricardo Garla^{54,55}, Kirk Gastrich³, Lachlan George², Tommaso Giarrizzo^{56,57}, Rory Graham⁵, Tristan L. Guttridge^{59,60}, Valerie Hagan¹³, Royale S. Hardenstine^{16,22}, Stephen M. Heck¹³, Aaron C. Henderson⁶¹, Patricia Heithaus³, Heidi Hertler⁶¹, Mauricio Hoyos Padilla^{62,63}, Robert E. Hueter^{64,65}, Rima W. Jabado^{1,66}, Jean-Christophe Joyeux⁶⁷, Vanessa Jaiteh^{68,69}, Mohini Johnson⁷⁰, Stacy D. Jupiter⁷¹, Muslimin Kaimuddin^{72,70}, Devanshi Kasana³, Megan Kelley³, Steven T. Kessel⁷³, Benedict Kiilu⁷⁴, Taratau Kirata⁷⁵†, Baraka Kuguru⁷⁶, Fabian Kyne⁷⁷, Tim Langlois^{78,79}, Frida Lara^{80,81}, Jaedon Lawe⁸², Elodie J. I. Lédée¹, Steve Lindfield⁸³, Andrea Luna-Acosta⁸⁴, Jade O. Maggs⁸⁵, B. Mabel Manjaji-Matsumoto⁸⁶, Andrea Marshall⁸⁷, Lucy Martin⁸⁸, Daniel Mateos-Molina^{89,90}, Philip Matich⁶⁰, Erin McCombs⁹¹, Ashlie McIvor^{22,92}, Dianne McLean^{6,93}, Llewelyn Meggs⁸², Stephen Moore¹, Sushmita Mukherji^{1,2}, Ryan Murray⁹⁴, Stephen J. Newman⁹⁵, Josep Nogués⁸⁸, Clay Obota^{96,97}, Domingo Ochavillo⁹⁸, Owen O'Shea^{99,100}, Kennedy E. Osuka^{96,101}, Yannis P. Papastamatiou³, Nishan Perera²³, Bradley Peterson¹⁴, Caio R. Pimentel^{67,102}, Fabián Pina-Amargós^{103,104}, Hudson T. Pinheiro¹⁰⁵, Alessandro Ponzo¹⁰⁶, Andhika Prasetyo¹⁰⁷, L. M. Sjamsul Quamar¹⁰⁸, Jessica R. Quinlan³, José Amorim Reis-Filho¹⁰⁹, Hector Ruiz¹¹⁰, Alexei Ruiz-Abierno¹⁰⁴, Enric Sala¹¹¹, Pelayo Salinas-de-León^{112,113}, Melita A. Samoilys^{96,114}, William R. Sample³, Michelle Schärer-Umpierre¹¹⁰, Audrey M. Schlaff¹, Kurt Schmid^{55,115}, Sara N. Schoen³, Nikola Simpson¹¹⁶, Adam N. H. Smith¹¹⁷, Julia L. Y. Spaet¹¹⁸, Lauren Sparks¹¹⁹, Twan Stoffers¹²⁰, Akshay Tanna²⁴, Rubén Torres¹²¹, Michael J. Travers⁹⁵, Maurits van Zinnicq Bergmann^{3,58}, Laurent Vigliola¹²², Juney Ward¹²³, Joseph D. Warren¹⁴, Alexandra M. Watts^{47,124}, Colin K. Wen¹²⁵, Elizabeth R. Whitman³, Aaron J. Wirsing¹²⁶, Aljoscha Wothke⁴⁰, Esteban Zarza-González^{127,128}. Demian D. Chapman^{3,60}

A global survey of coral reefs reveals that overfishing is driving resident shark species toward extinction, causing diversity deficits in reef elasmobranch (shark and ray) assemblages. Our specieslevel analysis revealed global declines of 60 to 73% for five common resident reef shark species and that individual shark species were not detected at 34 to 47% of surveyed reefs. As reefs become more shark-depleted, rays begin to dominate assemblages. Shark-dominated assemblages persist in wealthy nations with strong governance and in highly protected areas, whereas poverty, weak governance, and a lack of shark management are associated with depauperate assemblages mainly composed of rays. Without action to address these diversity deficits, loss of ecological function and ecosystem services will increasingly affect human communities.

oral reef ecosystems are under increasing pressure from human activitiesincluding intense fishing, degraded water quality, and climate change (1, 2)—that threaten species supporting a wide range of ecosystem functions (3). Sharks and rays (hereafter "elasmobranchs") have diverse roles

risk extinction categories by the International Union for the Conservation of Nature (IUCN).

With ~37% of all elasmobranch species threat-

on coral reefs as predators and prev across multiple trophic levels and in the cycling and movement of nutrients (3-5). Recent evidence indicates that overfishing has driven sharks toward functional extinction on many reefs. In a global survey, sharks were not observed on nearly 20% of reefs surveyed (6). Yet until recently, reef shark species were listed in lower ened with extinction (7), a key question Check for coral reef ecosystems lies in understand the global extent of species loss in elasmobranch assemblages. We characterized elasmobranch assemblage structure on coral reefs across a gradient of human pressures to estimate the local depletion and global extinction risk of the most common reef species, revealing the human and environmental factors that influence assemblage structure and that lead to a deficit in predator diversity that could

affect reef ecological functioning.

To understand the extent of the reef elasmobranch diversity deficit, we surveyed 391 coral reefs in 67 nations and territories using 22,756 baited remote underwater video stations (BRUVS). We examined reef-level species richness, species composition of elasmobranch assemblages, and species relative abundance (MaxN; the maximum number of each species observed in a single frame of each 60-min deployment then averaged across all deployments on one reef) (8). We examined how elasmobranch species assemblages changed in response to human pressures, using unweighted pair group with arithmetic mean (UPGMA) clustering to identify reefs with the most similar assemblages (8). We then compared these clusters with estimated depletion of key resident elasmobranch species at the reef level and examined whether socioeconomic, management, or environmental factors could predict cluster membership, using linear discriminant analysis. Reef-level depletion was estimated by dividing the observed mean MaxN of a species at individual reefs by a model-estimated baseline abundance (without human pressures) for each sampling site (a small group of closely associated reefs) and subtracting this value from 1. Baseline abundance (also expressed as MaxN) was estimated from a general linear model relating observed MaxN to sampling site, human pressure [represented by total market gravity, the size and travel time to human markets (2)], and marine protected area (MPA) status [closed to all fishing, open to fishing, or restricted (some fishing but with restrictions)]. The baseline was estimated by setting all parameters to those expected at a site with no human pressure (gravity to the minimum for an ocean basin and protection status to closed) (8).

Sampling identified 104 distinct elasmobranch species or species complexes (table S1), representing more than 77% of elasmobranch species known to occur on coral reefs at some point during their lives (9). More than half (n = 53) of the species were rarely observed, with 10 or fewer sightings. We estimated reeflevel depletion for the nine most commonly occurring species of shark [n = 5; Caribbean]reef sharks (Carcharhinus perezi) and nurse sharks (Ginglymostoma cirratum) in the Atlantic; grey reef sharks (Carcharhinus amblyrhynchos),

Affiliations are listed at the end of this paper.

^{*}Corresponding author, Email: colin.simpfendorfer@icu.edu.au

blacktip reef sharks (Carcharhinus melanopterus), and whitetip reef sharks (Triaenodon obesus) in the Indo-Pacific] and rays [n = 4]; vellow stingrays (Urobatis jamaicensis) and southern stingrays (Hypanus americanus) in the Atlantic; blue spotted mask rays (Neotrygon spp.) and blue spotted ribbontail rays (Taeniura lymma and Taeniura lessoni) in the Indo-Pacific]. The Galapagos shark was excluded from estimates of global depletion because sampling only covered a relatively small proportion of its range, but the results for this species were broadly similar. The nine key resident species represented 77.7% of all elasmobranchs observed in the study and are those that serve important ecological roles (10) and contribute the most to, and underpin, livelihoods through fishing (II) and dive tourism (12).

We found that mean depletion of five key resident reef sharks on individual reefs ranged from 100% depletion (none observed) to 0% (no depletion), averaging 62.8% (Fig. 1A). Mean depletion of key resident reef sharks followed the overall decline in elasmobranch abundance as measured with MaxN (Fig. 1B), decreased as the fraction of the elasmobranch assemblage comprised of sharks decreased (Fig. 1C), and showed little change across a range of elasmobranch species richness (Fig. 1D); these patterns were generally consistent between ocean basins. Across the range of depletion, five main clusters of reefs were identified in the Atlantic, and eight were identified in the Indo-Pacific (Figs. 2 and 3), including at least one cluster in each ocean basin (cluster 1 in the Atlantic and cluster 2 in the Indo-Pacific) having shark populations in a relatively intact state, with low levels of depletion of the five main resident reef shark species (Caribbean reef and nurse sharks in the Atlantic; grey reef, blacktip reef, and whitetip reef sharks in the Indo-Pacific) (8). Remaining clusters represented assemblages with increasing depletion of resident shark species and greater proportions of the overall elasmobranch assemblage represented by rays (Figs. 2C and 3B). Both ocean basins show a similar transition through these assemblages as key resident shark species became depleted. The four key ray species (yellow and southern stingrays in the Atlantic; blue spotted mask and blue spotted ribbontail rays in the Indo-Pacific) increased only with depletion of one or more resident reef shark species, with rays dominating in the most shark-depleted areas. These predictable changes in assemblage provide the ability to infer the status of reef shark populations, and the level of human pressure they are experiencing, in future

Elasmobranch species assemblage clusters on reefs in both basins were significantly related to certain socioeconomic and manage-

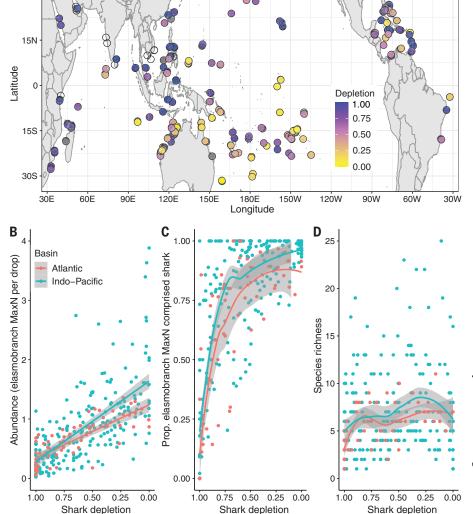
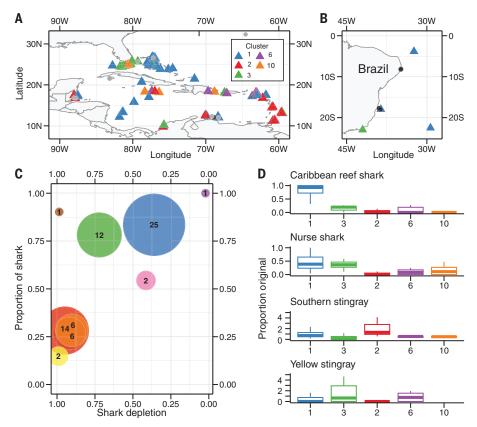


Fig. 1. The global decline of coral reef elasmobranchs. (**A**) Reef-scale estimates of depletion of resident coral reef shark species. Depletion is proportion of unfished population lost, represented as the measured MaxN as a proportion of MaxN in an unfished state (gravity, lowest in basin; MPA status, closed) (8). Open circles indicate no sharks or rays were observed; gray circles indicate none of the resident shark species used to calculate mean depletion were present. (**B**) Relationship between depletion of resident shark species and MaxN by ocean basin. (**C**) Relationship between depletion of resident shark species and the proportion of elasmobranch MaxN that comprised shark, demonstrating the transition from shark- to ray-dominated assemblages. (**D**) Relationship between depletion of resident shark species and species richness.

ment factors, with linear discriminant analysis (LDA) accounting for ~85% of variance between clusters (tables S2 and S3). Important socioeconomic factors included the Human Development Index (an index of a nation's level of education, life expectancy and standard of living) and Voice and Accountability Index (an index of the extent to which people in each nation can participate in governance, free expression, free media, and free association). Important management factors were whether the reef occurred in a marine pro-

tected area (MPA) or whether a reef was within a nation where all targeted shark fishing and trade is prohibited, known as a "shark sanctuary." Given that shark sanctuaries have largely been implemented in nations in which fishing for sharks was limited for economic or cultural reasons (6), their effectiveness as tools for recovering reef shark populations remains an open question. Total market gravity was more important in the Indo-Pacific than the Atlantic, possibly because remote reefs (>4 hours travel time from human settlements)

Fig. 2. Structure of shark and ray assemblages on Atlantic coral reefs. (A and B) Clusters of reefs with similar species composition from UPGMA clustering of 106 reefs in the Atlantic basin based on a global set of 31 coral reefassociated species. Five main clusters, representing 87.0% of reefs, were identified. Their locations are indicated with colored triangles. Reefs with minor clusters are indicated with gray dots (n = 7). Reefs where no elasmobranchs were observed are indicated with black dots (n = 5). (C) Regime plot showing all species assemblage clusters as a function of the mean depletion of the resident reef shark species (Caribbean reef and nurse sharks) and the proportion of all observed elasmobranchs that were sharks. Size of points (and numbers) indicate the number of reefs in each cluster, and colors indicate cluster identity as per (A). (D) Population level relative to original levels of four resident reef species in each of the five main clusters. Proportion of original level = 1 - depletion. Horizontal lines indicate mean, boxes indicates 25 to 75 percentile, and whiskers indicate 95% confidence interval.



are relatively rare in the Atlantic compared with the Indo-Pacific (fig. S1) (13). Environmental factors (coral cover and relief) had little influence in predicting cluster membership. Elasmobranch assemblage structure on coral reefs in both the Atlantic and Indo-Pacific are therefore mainly driven by management and socioeconomic factors, with shark-dominated assemblages more likely to occur in wealthy, well-governed nations and in highly protected areas or shark sanctuaries, whereas poverty, limited governance, and a lack of shark protection are associated with assemblages mainly composed of rays.

To further characterize the diversity deficits that underpin these assemblage differences, we compared species observations in our BRUVS with their historical ranges drawn from published literature, including historical accounts. and found that sharks were not detected at 13.6% of reefs (19 Atlantic and 34 Indo-Pacific), whereas ravs were not detected at 21.5% of reefs (10 Atlantic and 74 Indo-Pacific); both groups were not detected at 6.6% of reefs surveved (5 Atlantic and 19 Indo-Pacific). At the species level, absences were severe. On the basis of their known historic distribution, deficits were 46.9% of reefs (112 of 246) for blacktip reef sharks, 41.3% (31 of 75) for Caribbean reef sharks, 40.8% (102 of 250) for grey reef sharks, 36.2% (89 of 246) for whitetip reef sharks, and 34.7% (n = 26 of 75) for nurse sharks (fig. S2). Among rays, deficits were even more stark: 78.9% (75 of 95) for yellow stingray, 62.8% (81 of 129) for blue spotted ribbontail rays, and 55.6% (79 of 142) for blue spotted maskrays. An exception was the southern stingray, which was not detected at only 19.8% (n =20 of 101) of expected reefs in the Atlantic. A failure to detect rays may not always indicate absence because they are often cryptic and therefore missed on BRUVS, especially when sharks are present (14). Collectively, these diversity deficits show that elasmobranch loss on coral reefs is more extensive than previously demonstrated, with widespread losses of key species across many of the world's coral reefs, especially in Asia, eastern Africa, continental South America, and the central-eastern Caribbean.

Previous estimates of the status of reef shark and ray species have been geographically limited, varying among surveyed reefs from very high abundances (15) to local extinction (16). This disparity has made it difficult to assess the global status of individual species. Therefore, we used our estimates of reef-level depletion to estimate the global depletion and extinction risk of the most common resident reef sharks (five species) and rays (four species). Mean and standard error reef-level depletion was calculated within jurisdictions (nations or remote territories) and used to produce confidence intervals for jurisdic-

tional depletion levels. To estimate an overall global depletion level by species, we weighted the jurisdictional depletion by the percentage of the world's coral reefs in their waters and produced a weighted global mean depletion (8). Extinction risk was estimated by comparing proportional global depletion to the criteria for the IUCN Red List A2 (population decline) category (17), assuming that the decline had occurred in the past three generations (29 to 90 years). In IUCN assessments before the availability of this global survey, all reef-resident shark species were considered at lower risk of extinction (Near Threatened) (18). Grey reef shark had the highest level of global decline [69.8% ± 1 standard error (SE) 62.6 to 77.1], followed by nurse shark ($68.6\% \pm 49.7$ to 87.4), Caribbean reef shark (64.8% ± 42.0 to 87.5), blacktip reef shark (64.5% \pm 58.7 to 70.4), and whitetip reef shark ($60.4\% \pm 51.2$ to 70.2) (Fig. 4). The estimated declines of resident species of reef sharks met the IUCN Red List criteria for Endangered. Population changes of rays were more variable, with increasing populations in some nations and declines in others (fig. S3), reflecting the compositional changes seen across our gradient of human pressures. When examined at the global level, no ray species examined met criteria for elevated extinction risk, which is consistent with current nonthreatened status of these species on the Red List.

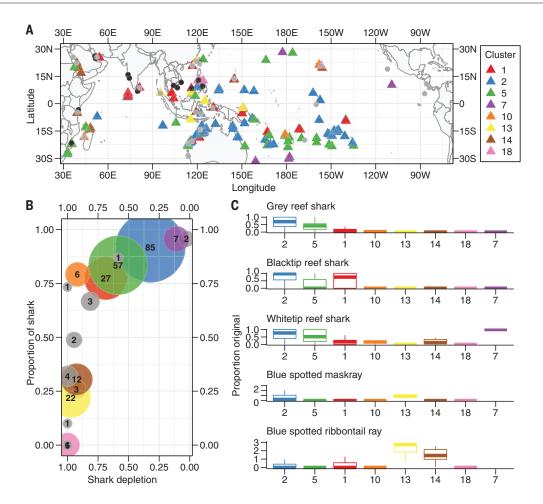


Fig. 3. Structure of shark and ray assemblages on Indo-Pacific coral reefs. (A) Clusters of reefs with similar species composition from UPGMA clustering of 285 reefs in the Indo-Pacific basin based on a global set of 31 coral reefassociated species. Eight main clusters, representing 82.1% of reefs, were identified. Their locations are indicated with colored triangles. Reefs with minor clusters are indicated with gray dots (n = 30). Reefs where no elasmobranchs were observed are indicated with black dots (n = 21). (B) Regime plot showing all species assemblage clusters as a function of the mean depletion of the

resident species of reef shark (grey reef, blacktip reef, whitetip reef, and Galapagos sharks) and the proportion of all observed elasmobranchs that were sharks. Size of points (and numbers) indicate the number of reefs in each cluster, and colors indicate cluster identity as per (A); minor clusters are indicated in gray. (C) Population level relative to original levels of five core shark and ray species in each of the eight main species assemblage clusters. Proportion of original level = 1 – depletion. Horizontal lines indicate mean, boxes indicate 25 to 75 percentile, and whiskers indicate 95% confidence interval.

Our study of nations hosting ~90% of global reefs reveals that resident reef shark species are at much higher risk of extinction than previously thought. Local declines, shaped by human pressures that vary across ocean basins, have led to consistent changes in the structure of coral reef elasmobranch assemblages that may have profound effects on the broader ecosystem. The direct and indirect effects of fishing have driven shifts in species composition from shark-dominated to ray-dominated assemblages and ultimately the complete loss of sharks and rays at a small proportion (~7%) of reefs surveyed. In addition to changes in the structure of assemblages, all major resident shark species have declined to such levels that they qualify as Endangered by the IUCN Red List Criteria. These changes wrought on coral reef elasmobranch assemblages demonstrate the pervasiveness of fishing on coral reefs (19) and the substantial risks to reef-dependent human communities of continued overfishing. Elasmobranch species vary widely in their economic value, with some fished for subsistence, others fished for local or export markets, and others valued alive as tourism resources (12, 20). Thus, understanding threats and conservation options for rebuilding populations at a species level will assist in developing effective management of coral reef elasmobranchs as part of a sustainable social-ecological system.

Although reef sharks are at considerable risk over broad spatial scales, our results show that declines at one reef will have little effect on reefs tens to hundreds of kilometers dis-

tant. Thus, despite populations being functionally extinct at the reef level, the potential to rebuild abundances remains relatively high if there are protected areas or strong fisheries management within a region (6). These source populations are present among many small oceanic islands where low human populations and the high cultural value of sharks has resulted in fishing levels that are below those seen elsewhere (21). MPAs also provide the opportunity to act as source populations; however, their designation alone is insufficient to deliver benefits. As others have observed (22), high compliance is required. We show that there are reefs in regions with widespread depletion of reef shark species that had metrics indicating that they are in a relatively healthy state compared with those around them. These

Fig. 4. Depletion of core coral reef shark species in the Indo-Pacific and Atlantic basins at national or near-national scale. (A) Indo-Pacific basin. (B) Atlantic basin. Depletion was calculated by comparing reef-level species MaxN values to unfished, estimated by using a linear model in which market gravity (a measure of the human pressure from population and access to reefs) was set to the ocean basin minimum and reef protected status was "closed" (no take MPA) (8). Reef-level depletion scores were modeled by nation and used to estimate a global level of depletion (vertical dashed lines) ± 1 standard error (shaded area) calculated by weighting national-level depletion by coral reef area (as a percent of global total coral reef area that occurs within the range of each shark species).

included Tubbataha (Philippines), Sipidan (Island Malaysia), Glover's Reef and Lighthouse Reef (Belize), and Misool (Indonesia); in all of these locations, there are programs to actively manage and enforce MPA regulations that are likely to account for these successes (23–25).

Multiple nations have strong management measures (such as spatial protections and/or fishing restrictions) in place that benefit reef species. This study builds the case that species-specific reef shark management provides the best way forward for conservation and rebuilding of reef sharks in places where they have declined, among nations with the desire and capacity to do so (7, 8). Recent studies show that populations of reef sharks can rebound in under a decade if appropriate management strategies that reduce fishing pressure are in place (26). Although direct management is critical, local and national socioeconomic factors that affect the ability of

nations to develop, implement, and enforce regulations, and the likelihood that fishers comply with regulations, will be critical to maintaining or rebuilding populations and diverse elasmobranch assemblages. If not addressed, pressures causing the shark and ray diversity deficits we outline will continue to result in a loss of species, ecological functions, and ecosystem services that support sustainable livelihoods for millions of people worldwide.

REFERENCES AND NOTES

- T. P. Hughes, D. R. Bellwood, S. R. Connolly, H. V. Cornell, R. H. Karlson, *Curr. Biol.* 24, 2946–2951 (2014).
- J. E. Cinner et al., Proc. Natl. Acad. Sci. U.S.A. 115, E6116–E6125 (2018).
- G. Roff et al., Trends Ecol. Evol. 31, 395–407 (2016).
- K. I. Flowers, M. R. Heithaus, Y. P. Papastamatiou, Fish Fish. 22, 105–127 (2021).
- J. J. Williams, Y. P. Papastamatiou, J. E. Caselle, D. Bradley, D. M. P. Jacoby, Mobile marine predators: an understudied source of nutrients to coral reefs in an unfished atoll. *Proc. R. Soc. B Biol. Sci.* 10.1098/rspb.2017.2456 (2018)
- 6. M. A. MacNeil et al., Nature 583, 801-806 (2020).
- 7. N. K. Dulvy et al., Curr. Biol. **31**, 4773–4787.e8
- Materials and methods are available as supplementary materials.
- 9. C. S. Sherman et al., Nat. Commun. 14, 15 (2023).
- M. R. Heupel, Y. P. Papastamatiou, M. Espinoza, M. E. Green,
 C. A. Simpfendorfer, Front. Mar. Sci. 6, 12 (2019).
- S. A. Appleyard, W. T. White, S. Vieira, B. Sabub, Sci. Rep. 8, 6693 (2018).
- A. M. Cisneros-Montemayor, M. Barnes-Mauthe,
 D. Al-Abdulrazzak, E. Navarro-Holm, U. R. Sumaila, *Oryx* 47, 381–388 (2013).
- 13. T. R. McClanahan, Mar. Policy 119, 104022 (2020).
- C. S. Sherman, M. R. Heupel, S. K. Moore, A. Chin,
 C. A. Simpfendorfer, Mar. Ecol. Prog. Ser. 641, 145–157 (2020)
- 15. J. Mourier et al., Curr. Biol. 26, 2011-2016 (2016)
- 16. M. O. Nadon et al., Conserv. Biol. 26, 493–503 (2012). 17. IUCN, IUCN Red List categories and criteria, version 3.1
- IUCN, IUCN Red List categories and criteria, version 3.. (IUCN, ed. 2, 2012).
- 18. N. K. Dulvy et al., Aquat. Conserv. 26, 134-153 (2016).
- 19. T. D. Eddy et al., One Earth 4, 1278-1285 (2021).
- H. Booth, D. Squires, E. J. Milner-Gulland, *Ocean Coast. Manage*. 182, 104994 (2019).
- 21. J. S. Goetze et al., PLOS ONE 13, e0200960 (2018).
- 22. G. J. Edgar et al., Nature 506, 216-220 (2014).
- 23. R. Murray et al., J. Asia-Pac. Biodivers. 12, 49-56 (2019).
- 24. V. F. Jaiteh et al., Front. Mar. Sci. 3, (2016). 25. G. Clementi et al., Mar. Ecol. Prog. Ser. 661, 175–186
- (2021). 26. C. W. Speed, M. Cappo, M. G. Meekan, *Biol. Conserv.* **220**,
- C. W. Speed, M. Cappo, M. G. Meekan, *Biol. Conserv.* 220 308–319 (2018).
- C. Simpfendorfer, Widespread diversity deficits of coral reef sharks and rays. Dryad (2023); https://doi.org/10.5061/ dryad.qbzkh18h0.
- C. Simpfendorfer, Widespread diversity deficits of coral reef sharks and rays. Zenodo (2023); https://doi.org/10.5281/ zenodo.7030578.

ACKNOWLEDGMENTS

We thank our individual funders for country-specific deployments, whose contributions greatly enhanced the sampling coverage of the projects; all of the government permitting agencies that allowed us to work in their waters; and the Global FinPrint volunteers from Stony Brook University, Florida International University, James Cook University, the Aquarium of the Pacific, and Shedd Aquarium who watched the BRUVS footage. Funding: Core funding for Global FinPrint was provided by the Paul G. Allen Family Foundation (to D.D.C. and M.R.Hei.). Author contributions: Conceptualization: D.D.C., M.R.Hei., C.A.S., M.R.Heu., M.A.M., M.M., and E.H. Methodology: D.D.C., M.R.Hei., C.A.S., M.R.Heu., M.A.M., M.M., and E.H. Investigation: All authors. Visualization: C.A.S. Funding acquisition: D.D.C. and M.R.Hei. Project administration: D.D.C., M.R.Hei., C.A.S., M.R.Heu., M.A.M., M.M., and E.H. Writing original draft: C.A.S., D.D.C., M.R.Hei., M.R.Heu., M.A.M., M.M., E.H., and C.S.S. Writing - review and editing: All authors. Competing interests: The authors declare that they have no competing interests Data and materials availability: Data files have been deposited in Dryad (27), and R script has been deposited in Zenodo (28). License information: Copyright © 2023 the authors, some rights reserved; exclusive licensee American Association for the Advancement of Science. No claim to original US government works. https://www. science.org/about/science-licenses-journal-article-reuse

¹College of Science and Engineering, James Cook University, Townsville, QLD, Australia. ²Institute for Marine and Antarctic Studies, University of Tasmania, Hobart, TAS, Australia. ³Institute of

Environment, Department of Biological Sciences, Florida International University, North Miami, FL, USA. 4Australian Institute of Marine Science, Townsville, OLD, Australia, 5Ocean Frontier Institute, Department of Biology, Dalhousie University, Halifax, NS, Canada. ⁶Australian Institute of Marine Science, Perth, WA, Australia. ⁷School of Molecular and Life Sciences, Curtin University, Bentley, WA, Australia. 8Earth to Ocean Group, Biological Sciences, Simon Frasei University, Burnaby, BC, Canada. 9School of Molecular and Life Sciences, Curtin University, Perth, WA, Australia. 10 Marine Science Program, Biodiversity and Conservation Science, Department of Biodiversity, Conservation and Attractions, Perth. WA, Australia. ¹¹Centre for Sustainable Ecosystems Solutions, School of Earth, Atmospheric and Life Sciences, University of Wollongong, Wollongong, NSW, Australia. 12 Australian Institute of Marine Science, Darwin, NT, Australia. ¹³Sharks and Rays Conservation Program, Mote Marine Laboratory, Sarasota, FL, USA. 14School of Marine and Atmospheric Sciences, Stony Brook University, Stony Brook, NY, USA. ¹⁵International Pole and Line Foundation-Maldives, Malé, Republic of Maldives. 16 Maldives Marine Research Institute, Ministry of Fisheries, Marine Resources and Agriculture, Malé, Republic of Maldives. ¹⁷Red Sea Global, Department of Environmental Protection and Regeneration, AlRaidah Digital City, Riyadh, Saudi Arabia. ¹⁸Kap Natirel NGO, Fort l'Olive, Guadeloupe, France. ¹⁹Mahonia Na Dari Research and Conservation Centre, Kimbe, Papua New Guinea. 20 South African Institute for Aquatic Biodiversity, National Research Foundation, Makhanda, South Africa. 21 Department of Zoology and Entomology, Rhodes University, Makhanda, South Africa. ²²Red Sea Research Center, King Abdullah University of Science and Technology, Thuwal, Saudi Arabia. ²³Marine Conservation, Madagascar Program, Wildlife Conservation Society, Antananarivo, Madagascar. 24Blue Resources Trust, Colombo, Sri Lanka. ²⁵Cape Eleuthera Institute, Cape Eleuthera, Eleuthera, The Bahamas. ²⁶Center for Sustainable Development, College of Arts and Sciences, Oatar University, Doha, Oatar, 27Georgia Aquarium-IUCN Center for Species Survival, Atlanta, GA, USA. ²⁸School of Natural Sciences, Macquarie University, Sydney, NSW, Australia. ²⁹Bimini Biological Field Station, Bimini, Bahama. ³⁰Centro de Investigación en Ciencias del Mar y Limnología, Universidad de Costa Rica, San José, Costa Rica. 31 MigraMar, Olema, CA, USA. ³²Department of Biology, Albion College, Albion, MI, USA. ³³Marine Science Institute, University of California Santa Barbara, Santa Barbara, CA, USA. 34Coastal Impact, Goa, India. 35College of Arts, Society, and Education, James Cook University, Townsville, QLD, Australia. 36Centre Universitaire de Formation et de Recherche de Mayotte, Dembeni, France. ³⁷Paris Sciences Lettres, Centre de Recherche Insulaire et Observatoire de l'Environnement, Opunohu Bay, Papetoai, French Polynesia, 38Laboratoires d'Excellence Corail, Ecole Pratique des Hautes Etudes, Perpignan, France. 39School of Biosciences, Cardiff University, Cardiff, UK. 40 Environmental Research Institute Charlotteville, Charlotteville, Trinidad and Tobago. 41 Australian Research Council Centre of Excellence for Coral Reef Studies, James Cook University, Townsville, QLD, Australia. 42Sharks Pacific, Rarotonga, Cook Islands. 43 Oceans and Atmosphere, Commonwealth Scientific and Industrial Research Organization, Hobart, TAS, Australia. 44Wageningen Marine Research, Wageningen University & Research, IJmuiden, Netherlands. 45Graduate School for Global Environmental Studies, Sophia University, Tokyo, Japan. 46Waitt Institute, La Jolla, CA, USA. ⁴⁷Reef Systems Ecology and Conservation Lab, Departamento de Biologia Marinha, Universidade Federal Fluminense, Rio de Janeiro. Brazil, 48 Marine Megafauna Foundation, Palm Beach, FL, USA, ⁴⁹Oceanographic Research Institute, Durban, South Africa. ⁵⁰TRAFFIC International, Cambridge, UK. 51 College of Arts, Science, and Education, Florida International University, North Miami, FL, USA. 52 Science Department, Georgia Jones-Ayers Middle School, Miami, FL, USA. ⁵³Division of Aquatic Resources, Department of Land and Natural Resources, Honolulu, HI, USA. 54 Centro de Biociências, Departmento de Botânica e Zoologia, Universidade Federal do Rio Grande do Norte, Brazil. 55Beacon Development Company, King Abdullah University of Science and Technology, Thuwal, Saudi Arabia. ⁵⁶Instituto de Ciencias do Mar, Universidade Federal do Ceará, Fortaleza, Brazil. 57 Grupo de Ecologia Aquática, Espaço Inovação do Parque de Ciência e Tecnologia Guamá, Guamá, Pará, Brazil. 58 Independent consultant, Hull, UK. ⁵⁹Bimini Biological Field Station Foundation, South Bimini, The Bahamas. 60 Saving the Blue, Cooper City, FL, USA. 61 The School for Field Studies, Center for Marine Resource Studies, South Caicos, Turks and Caicos Islands. ⁶²Pelagios Kakunjá, La Paz, Mexico. ⁶³Fins Attached, Colorado Springs, CO, USA. 64Center for Shark Research, Mote Marine Laboratory, Sarasota, FL, USA. ⁶⁵OCEARCH, Park City, UT, USA. ⁶⁶Elasmo Project, Dubai, United Arab Emirates. ⁶⁷Departamento de Oceanografia e Ecologia, Universidade Federal do Espírito Santo, Vitória, Espírito Santo, Brazil. 68 Murdoch University, Murdoch, WA, Australia. 69 Centre for Development and Environment, University of

Bern, Bern, Switzerland. 70 Operation Wallacea, Spilsby, Lincolnshire,

UK. 71 Melanesia Program, Wildlife Conservation Society, Suva, Fiji.

⁷²Wasage Divers, Wakatobi & Buton, Southeast Sulawesi, Indonesia. ⁷³Daniel P. Haerther Center for Conservation and Research, John G. Shedd Aquarium, Chicago, IL, USA. 74Kenya Fisheries Service, Mombasa, Kenya. ⁷⁵Ministry of Fisheries and Marine Resources, Kiritimati, Kiribati. ⁷⁶Tanzania Fisheries Research Institute, Dar Es Salaam, Tanzania. ⁷⁷University of the West Indies, Kingston, Jamaica. ⁷⁸School of Biological Sciences, University of Western Australia, Perth, WA, Australia. 79The UWA Oceans Institute, University of Western Australia, Perth, WA, Australia. 80 Departamento de Pesquerias, Centro Interdisciplinario de Ciencias Marinas del IPN, La Paz, Baja California Sur, Mexico. 81 Pelagios Kakunjá, La Paz, Baja California Sur, Mexico. ⁸²Yardie Environmental Conservationists Limited, Kingston, Jamaica. ⁸³Coral Reef Research Foundation, Koror, Palau. ⁸⁴Departamento de Ecología y Territorio, Facultad de Estudios Ambientales y Rurales, Pontificia Universidad Javeriana, Bogotá, Colombia. 85 National Institute of Water and Atmospheric Research, Auckland, New Zealand, 86Borneo Marine Research Institute, Universiti Malaysia Sabah, Kota Kinabalu, Sabah, Malaysia. 87 Marine Megafauna Foundation, West Palm, FL, USA. ³⁸Island Conservation Society Seychelles, Victoria, Mahé, Seychelles. ⁸⁹Emirates Nature - World Wide Fund for Nature, Dubai, United Arab Emirates. 90 College of Marine Sciences and Aquatic Biology, University of Khorfakkan, Sharjah, UAE. 91 Aquarium of the Pacific, Long Beach, CA, USA. 92 Marine and Environmental Sciences Centre/ Aquatic Research Network, Regional Agency for the Development of Research, Technology and Innovation, Funchal, Madeira, Portugal. 93Oceans Institute, University of Western Australia, Perth, WA, Australia. 94Inland Fisheries Ireland, Dublin, Ireland. 95Western Australian Fisheries and Marine Research Laboratories, Department of Primary Industries and Regional Development, Government of Western Australia, Hillarys, WA, Australia. 96CORDIO East Africa, Mombasa, Kenya. 97Blue Ventures, Mombasa, Kenya. 98 American Samoa Department of Marine and Wildlife Resources, Pago Pago, American Samoa. 99The Centre for Ocean Research and Education, Gregory Town, Eleuthera, The Bahamas. 100 Department of Ocean Science, Memorial University, NL, Canada. ¹⁰¹Department of Environment and Geography, University of York, York, UK. ¹⁰²Departamento de Ciências Agrárias e Biológicas, Universidade Federal do Espírito Santo, São Mateus, Espírito Santo, Brazil. 103 Blue Sanctuary-Avalon, Jardines de la Reina, Cuba. 104 Centro de Investigaciones Marinas, Universidad de La Habana, Habana, Cuba. ⁵Center for Marine Biology, University of São Paulo, São Sebastião, São Paulo, Brazil. 106Large Marine Vertebrates Research Institute Philippines, Puerto Princesa City, Palawan, Philippines. 107Center for Fisheries Research, Ministry for Marine Affairs and Fisheries, Jakarta Utara, Indonesia. ¹⁰⁸Fisheries Department, Universitas Dayanu Ikhsanuddin, Bau Bau, Southeast Sulawesi, Indonesia. 109 Programa de Pós Graduação em Ecologia: Teoria, Aplicação e Valores, Instituto de Biologia, Universidade Federal da Bahia, Salvador, BA, Brazil. ¹¹⁰HJR Reefscaping, Boquerón, Puerto Rico. ¹¹¹Pristine Seas, National Geographic Society, Washington, DC, USA. 112 Charles Darwin Research Station, Charles Darwin Foundation, Puerto Ayora, Galapagos Islands, Ecuador. 113 Save Our Seas Foundation Shark Research Center and Guy Harvey Research Institute, Nova Southeastern University, Dania Beach, FL, USA. ¹¹⁴School of Pure and Applied Sciences, Pwani University, Kilifi, Kenya. 115Thurgau Hunting and Fishing Administration, Frauenfeld, Switzerland. ¹¹⁶SalvageBlue, Kingstown, Saint Vincent and the Grenadines. 117School of Mathematical and Computational Sciences, Massey University, Auckland, New Zealand. 118 Evolutionary Ecology Group, Department of Zoology, University of Cambridge, Cambridge, UK. ¹¹⁹Indo Ocean Project, Jln Toyapakeh DESA Toyapakeh, Nusa Penida, Bali, Indonesia. ¹²⁰Aquaculture and Fisheries Group, Wageningen University & Research, Wageningen, Netherlands. 121 Reef Check Dominican Republic, Santo Domingo, Dominican Republic. 122 Institut de Recherche pour le Développement, UMR Entropie (IRD-UR-UNC-CNRS-IFREMER), Nouméa, New Caledonia, France. ¹²³Secretariat of the Pacific Regional Environment Programme, Apia, Samoa. 124 Department of Natural Sciences, Faculty of Science Engineering, Manchester Metropolitan University, Manchester, UK. ¹²⁵Department of Life Science, Tunghai University, Taichung, Taiwan. 126 School of Environmental and Forest Sciences, University of Washington, Seattle, WA, USA. 127 GIBEAM Research Group, Universidad del Sinú, Cartagena, Colombia. 128 Corales del Rosario and San Bernardo National Natural Park Colombia

SUPPLEMENTARY MATERIALS

science.org/doi/10.1126/science.ade4884 Materials and Methods Figs. S1 to S8 Tables S1 to S9 References (29–37) MDAR Reproducibility Checklist

Submitted 31 August 2022; accepted 27 April 2023 10.1126/science.ade4884



Widespread diversity deficits of coral reef sharks and rays

Colin A. Simpfendorfer, Michael R. Heithaus, Michaelle R. Heupel, M. Aaron MacNeil, Mark Meekan, Euan Harvey, C. Samantha Sherman, Leanne M. Currey-Randall, Jordan S. Goetze, Jeremy J. Kiszka, Matthew J. Rees, Conrad W. Speed, Vinay Udyawer, Mark E. Bond, Kathryn I. Flowers, Gina M. Clementi, Jasmine Valentin-Albanese, M. Shiham Adam, Khadeeja Ali, Jacob Asher, Eva Aylagas, Ocane Beaufort, Cecilie Benjamin, Anthony T. F. Bernard, Michael L. Berumen, Stacy Bierwagen, Chico Birrell, Erika Bonnema, Rosalind M. K. Bown, Edward J. Brooks, J. Jed Brown, Dayne Buddo, Patrick J. Burke, Camila Cceres, Marta Cambra, Diego Cardeosa, Jeffrey C. Carrier, Sara Casareto, Jennifer E. Caselle, Venkatesh Charloo, Joshua E. Cinner, Thomas Claverie, Eric E. G. Clua, Jesse E. M. Cochran, Neil Cook, Jessica E. Cramp, Brooke M. DAlberto, Martin de Graaf, Mareike C. Dornhege, Mario Espinoza, Andy Estep, Lanya Fanovich, Naomi F. Farabaugh, Daniel Fernando, Carlos E. L. Ferreira, Candace Y. A. Fields, Anna L. Flam, Camilla Floros, Virginia Fourqurean, Laura Gajdzik, Laura Garca Barcia, Ricardo Garla, Kirk Gastrich, Lachlan George, Tommaso Giarrizzo, Rory Graham, Tristan L. Guttridge, Valerie Hagan, Royale S. Hardenstine, Stephen M. Heck, Aaron C. Henderson, Patricia Heithaus, Heidi Hertler, Mauricio Hoyos Padilla, Robert E. Hueter, Rima W. Jabado, Jean-Christophe Joyeux, Vanessa Jaiteh, Mohini Johnson, Stacy D. Jupiter, Muslimin Kaimuddin, Devanshi Kasana, Megan Kelley, Steven T. Kessel, Benedict Kiilu, Taratau Kirata, Baraka Kuguru, Fabian Kyne, Tim Langlois, Frida Lara, Jaedon Lawe, Elodie J. I. Lde, Steve Lindfield, Andrea Luna-Acosta, Jade Q. Maggs, B. Mabel Manjaji-Matsumoto, Andrea Marshall, Lucy Martin, Daniel Mateos-Molina, Philip Matich, Erin McCombs, Ashlie McIvor, Dianne McLean, Llewelyn Meggs, Stephen Moore, Sushmita Mukherji, Ryan Murray, Stephen J. Newman, Josep Nogus, Clay Obota, Domingo Ochavillo, Owen O'Shea, Kennedy E. Osuka, Yannis P. Papastamatiou, Nishan Perera, Bradley Peterson, Caio R. Pimentel, Fabin Pina-Amargs, Hudson T. Pinheiro, Alessandro Ponzo, Andhika Prasetyo, L. M. Sjamsul Quamar, Jessica R. Quinlan, Jos Amorim Reis-Filho, Hector Ruiz, Alexei Ruiz-Abierno, Enric Sala, Pelayo Salinas de-Len, Melita A. Samoilys, William R. Sample, Michelle Schrer-Umpierre, Audrey M. Schlaff, Kurt Schmid, Sara N. Schoen, Nikola Simpson, Adam N. H. Smith, Julia L. Y. Spaet, Lauren Sparks, Twan Stoffers, Akshay Tanna, Rubn Torres, Michael J. Travers, Maurits van Zinnicq Bergmann, Laurent Vigliola, Juney Ward, Joseph D. Warren, Alexandra M. Watts, Colin K. Wen, Elizabeth R. Whitman, Aaron J. Wirsing, Aljoscha Wothke, Esteban Zarza-Gonzlez, and Demian D. Chapman

Science, 380 (6650), .

DOI: 10.1126/science.ade4884

Editor's summary

In recent years, much attention has been given to catastrophic declines in sharks. Most of this attention has focused on large pelagic species that are highly threatened by direct and indirect harvest. Simpfendorfer *et al.* looked globally at the smaller, coral reef–associated species of sharks and rays and found steep declines in shark species (see the Perspective by Shiffman). Five of the most common reef shark species have experienced a decline of up to 73%. As shark species decline on coral reefs, ray species increase, indicating a community-wide shift. Species are best protected when active protections are in place, suggesting routes for better conservation. —Sacha Vignieri

View the article online

https://www.science.org/doi/10.1126/science.ade4884 Permissions

https://www.science.org/help/reprints-and-permissions

Use of this article is subject to the Terms of service